

1 Combatting global grassland degradation

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25

26 Abstract

27 The World's grasslands are under severe threat from on-going degradation, yet they are
 28 largely ignored in sustainable development agendas. This degradation is undermining the
 29 capacity of grasslands to support biodiversity, ecosystem services, and human wellbeing. In
 30 this Perspective, we examine the current state of grasslands worldwide and explore the extent
 31 and dominant drivers of global grassland degradation. We identify actions that are critical to
 32 the development of socio-ecological solutions to combat degradation and promote restoration
 33 of global grasslands. Specifically, we argue that progress can be made by: increasing
 34 recognition of grasslands in global policy, developing standardised indicators of grassland
 35 degradation, using scientific innovation for effective restoration at regional and landscape
 36 scales, and enhancing knowledge transfer and data sharing on restoration experiences. The
 37 integration of these strategies into sustainability policy should help to halt grassland

38 degradation and enhance restoration success, and protect the socio-economic, cultural and
39 ecological benefits that grasslands provide.

40

41 Grasslands, comprising open grassland, grassy shrublands and savannah, cover about
42 40% of the Earth's surface and some 69% of the world's agricultural land area¹⁻³. Not only do
43 they serve as an important global reservoir of biodiversity, including many iconic and
44 endemic species, but also, they provide a wide range of material and non-material benefits to
45 humans and our quality of life. These benefits include a wide range of ecosystem services,
46 such as food production, water supply and regulation, carbon storage and climate mitigation,
47 pollination, and a host of cultural services¹⁻³. Despite its importance, grassland degradation is
48 widespread and accelerating in many parts of the world⁴⁻⁶ with as much as 49% of grassland
49 area worldwide having been degraded to some extent^{5,7,8}.

50 Grassland degradation poses an enormous threat to the hundreds of millions of people
51 who rely on grasslands worldwide for food, fuel, fibre and medicinal products, as well as
52 their multiple cultural values^{9,10}. In terms of livestock production, the global cost of grassland
53 degradation has been estimated at \$6.8 billion¹¹, with the impact on human welfare being
54 particularly severe in regions where most the population is below the poverty line. Grassland
55 degradation also creates major environmental problems, given that grasslands play a critical
56 role in biodiversity conservation, climate and water regulation, and global biogeochemical
57 cycles^{2,4}. For example, the conversion of tropical grassy biomes to arable cropland poses a
58 significant threat to biodiversity given that they have a vertebrate species richness
59 comparable to forests¹², while the recent expansion of croplands in United States has caused
60 widespread conversion of prairie grasslands, with considerable cost to wildlife⁶. Moreover,
61 the conversion of grasslands to arable cropland and disturbance through overgrazing, fire and
62 invasive species can lead to significant soil carbon loss¹³. Because of these problems,
63 grassland degradation represents a major global challenge that must be addressed if we are to
64 achieve key targets of biodiversity agendas, such as the Aichi Biodiversity Targets¹⁴ of the
65 Convention on Biological Diversity (CBD) and the United Nations Sustainable Development
66 Goals (SDGs), including restoration and sustainable use of terrestrial ecosystems, hunger and
67 poverty alleviation, and climate change mitigation. Combating degradation is also central to
68 the UN Decade on Ecosystem Restoration (2021-2030), which puts a firm focus on the urgent
69 need for strategies for the restoration of degraded ecosystems, including grasslands.

70 There are multiple, co-occurring drivers of grassland degradation, including over-
71 grazing, eutrophication, land conversion to forestry and crops, land abandonment, invasive

72 species, climate extremes and altered fire regimes (Figure 1). These drivers are also often
73 closely allied to socio-economic drivers, such as spatial expansion of humans, the economics
74 of land use, changes in affluence and dietary preferences, and in land tenure and a lack of
75 enforcement of land use rights. Climate change is also exacerbating grassland degradation,
76 especially due to fires and droughts, which are becoming more common and severe¹⁵. Despite
77 the importance of grasslands, there has been little progress in finding solutions to halting and
78 reversing global grassland degradation, which is compromising sustainable development and
79 the ecosystem services grasslands provide.

80 In this Perspective, we consider the current state of grasslands worldwide (comprising
81 open grassland, grassy shrublands, and savannah) and the ecosystem services they provide, and
82 explore what is known about the extent and dominant causes of grassland degradation. We then
83 consider the challenges that grassland restoration efforts face, from societal attitudes to
84 practical efforts, and identify actions that are critical to the development of socio-ecological
85 solutions to combat degradation and promote restoration of global grasslands. We argue that
86 progress can only be made through increasing recognition of grasslands in regional and global
87 policy, developing standardised indicators of degradation and restoration, introducing
88 ecological innovation into restoration, and enhancing knowledge transfer and data sharing
89 regarding restoration experiences. If adopted, these measures should benefit global policy
90 initiatives to tackle ecosystem degradation and restoration, and protect the many socio-
91 economic, cultural and ecological benefits that grasslands provide.

92

93 **Defining grasslands**

94

95 A key issue in the poor recognition of grasslands in policy is a lack of clarity in defining the
96 grassland biome. Grasslands can be defined in terms of both what they are, for example
97 dominated by grasses and graminoids, and what they are not, for example lacking extensive
98 trees or scrub. Broadly defined, grasslands include, amongst others, European meadows and
99 pastures, North American prairies, South American cerrados and pampas, African savannahs,
100 Australian rangelands, and Asian steppes. In many parts of the world and for many ecologists
101 and conservationists, ‘natural grasslands’ are the most important biome, and these are found in
102 both the tropical and temperate zones. Natural grasslands are generally found where the climate
103 and/or soils cannot support woody plant growth, such as above the tree-line or in the dry-lands
104 of the world. Often such factors interact with natural disturbance such as by fire or wild grazers,

105 which constrain growth of woody plants and maintain the grasses and associated low-growing
106 plants¹⁶. Indeed, changes in the balance of these interacting factors can trigger radical change.
107 For example, many parts of Africa (and elsewhere) are seeing ‘woody plant encroachment’
108 linked to a decline in burned area along with warmer and wetter climates¹⁷, while the Holocene
109 transition from extensive grasslands into the Sahara Desert was likely due to the onset of drier
110 conditions at the end of the African Humic Period¹⁸.

111 So-called ‘semi-natural grasslands’ are also important ecologically, and receive
112 special attention in Europe¹⁹, where they are important reservoirs for biodiversity²⁰. Such
113 grasslands have been produced by centuries of human activity, generally by traditional, low-
114 intensity agriculture involving livestock grazing, hay-cutting, and sometimes management
115 with fire. Semi-natural grasslands have been created by clearing natural vegetation such as
116 forest, but also by modifying natural grasslands. As agriculture has become increasingly
117 intensive over the last century, ‘improved’ grasslands of low plant species diversity have
118 come to dominate in many parts of the world, characterised by the use of inorganic fertilisers,
119 high stocking densities of productive livestock breeds, and often frequent re-sowing with
120 high-yielding plant varieties. In this Perspective, we focus on natural and semi-natural
121 grasslands, with improved grasslands generally considered a form of grassland degradation in
122 terms of loss of biodiversity and many ecosystem services¹⁹.

123 The vegetation of natural and semi-natural grasslands, while typically dominated by
124 grasses and graminoids, often has a large forb component and even scattered trees or scrub.
125 Indeed, in terms of species richness, forbs dominate grasslands worldwide, and are often the
126 most species-rich functional group, for example in North American prairie, African savannah,
127 Mongolian steppes, and South American pampas, campos and cerrados²¹. According to some
128 definitions, grasslands can also have a relatively high tree cover before they become classed as
129 forest²². Natural grasslands are often grazed by wild herbivores, including many large
130 mammals, but large areas are also used for domestic livestock. As such, meat, dairy and other
131 livestock products such as wool comprise the most prominent ecosystem services provided by
132 both natural and semi-natural grasslands. The livelihoods and cultures of some traditional
133 communities revolve almost entirely around their livestock systems^{20,23}, which in turn
134 maintains the diversity of livestock breeds²⁴.

135 Grasslands support a huge amount of biodiversity, hosting numerous species not found
136 in forests^{12,25} and distinct assemblages of native large herbivores and their predators, many of
137 which are threatened by extinction due to the combined impacts of hunting, encroachment by
138 humans and their livestock, and habitat loss²⁶. Many of these species are highly charismatic

139 and globally renowned, such as bison, wolves, rhinos and lions, and hence provide cultural
140 services, such as ecotourism, and related benefits²⁷. It is perhaps less acknowledged, however,
141 that grasslands provide many other ecosystem services, including holding flood waters,
142 providing clean water, regulating soil erosion, and providing wild food and medicines¹⁰. These
143 services and their importance vary geographically in relation to local environments and the
144 demands from people¹⁰. Cultural services include aesthetic value, hunting, heritage values and
145 resources for tourism and recreation, especially where grasslands define the ecology and
146 culture of an area²⁸.

147 The role of grasslands in carbon storage is often overlooked, although estimates suggest
148 that they contain 30% of the world's soil carbon stock²⁹ and that natural and sparsely grazed
149 grasslands act as a significant global carbon sink and provide an important climate cooling
150 service³⁰. However, there is a global trend of grasslands transitioning towards a net warming
151 effect on climate due to increased greenhouse gas emission associated with higher livestock
152 numbers and management intensification³⁰. Furthermore, natural and sparsely grazed
153 grasslands contain "irrecoverable carbon" that is vulnerable to land use conversion; once lost,
154 this carbon is not recoverable over timescales relevant to climate mitigation³¹. Nevertheless,
155 there is high potential for increasing soil carbon sequestration in grasslands via improved
156 grazing and by arresting grassland conversion and degradation³².

157

158

159 **Extent and drivers of degradation**

160

161 Globally, estimates suggest that as much as 49% of grassland area has been degraded to some
162 extent^{5,7,8}, although there is much uncertainty given that definitions of both grassland and
163 degradation vary. For example, improved grasslands can be conflated with the natural and
164 semi-natural grasslands considered here. The IPBES Assessment Report on Land Degradation
165 and Restoration³³ suggests that about half of the global natural grassland area is degraded to
166 some extent. More specific estimates include: about 90% of the UK's semi-natural grasslands
167 lost since the 1940s¹⁹, up to 90% of the vast grasslands of the Qinghai-Tibetan Plateau
168 degraded³⁴, and over 60% of the former grassland area of southern Brazil lost³⁵. As illustrated
169 by this geographic range, threats to natural and semi-natural grasslands are present across most
170 of the globe. However, current threat levels seem particularly high in the tropics¹⁶, while loss
171 of the small remaining areas of western European grassland has largely slowed³⁶. It is likely,
172 however, that ongoing environmental change will pose further threats to all grasslands. Climate

173 change has been shown in the past to cause grassland loss and degradation, and projected future
174 climate change will likely combine with human activities to cause increased woody plant
175 encroachment in some areas and desertification in others^{37,38}. An increasing and immediate
176 concern is the planned planting of trees on important grasslands across the globe, ostensibly to
177 meet afforestation targets for climate change mitigation³⁹.

178 There are many drivers of grassland degradation, which we classify for heuristic
179 purposes into three broad types: disturbances, anthropogenic inputs, and land use change. The
180 degree of disturbance is important; too little can drive a transition into more woody vegetation,
181 while too much can lead to soil erosion and desertification. In Europe, many semi-natural
182 grasslands were abandoned and lost to scrub in the 20th century as these low intensity grazing
183 systems became uneconomic⁴⁰. Over-grazing or a heightened fire frequency can disrupt
184 vegetation cover, cause soil erosion, and lead to desertification⁴¹. Considering the intensity of
185 anthropogenic inputs, natural and semi-natural grasslands tend to receive low-intensity
186 management, contrasting with the fertiliser inputs and re-sowing with productive cultivars to
187 support heavy grazing in improved grasslands. More general human inputs, such as through
188 atmospheric nitrogen deposition, can have similar effects to agricultural intensification⁴². The
189 third type comprises humans actively converting grasslands to other land uses, for instance to
190 arable farming, built infrastructure, and to forestry^{35,43}. A related problem is invasion by
191 introduced woody plants, such as *Prosopis juliflora*, which change the land cover⁴⁴.

192 These three drivers of grassland degradation interact and are also exacerbated by
193 climate change. For example, natural grasslands in the Americas, Australia and Africa are
194 being degraded due to woody plant encroachment, with the major causes thought to be a
195 combination of higher atmospheric CO₂ concentrations, reductions in fire or grazing pressures,
196 and warmer and wetter conditions⁴⁵. Conversely, in many parts of the world, such as Central
197 Asia and Africa, over-grazing combined with more intense and frequent droughts is
198 exacerbating grassland desertification and degradation⁴⁶.

199

200 **Halting degradation and promoting restoration**

201

202 Here, we propose five actions that should be deployed to develop effective socio-ecological
203 solutions to the degradation and restoration of global grasslands.

204

205 ***Policy recognition of grasslands.*** A key factor contributing to the degradation of grasslands is

206 the lack of representation of grassland in national and international policy. In fact, grasslands
207 are in many ways the ‘Cinderella’ of the world’s biomes, with the scientific and political focus
208 of global sustainability policy being on ecosystems such as oceans, freshwaters, forests and
209 croplands. Indeed, despite suffering some of the highest rates of degradation of any biome,
210 Dudley et al⁴⁷ reported that only 8% of grassland and savannah biomes protected. As such,
211 there is a pressing need for national and global policy to recognize the role of grasslands in
212 addressing climate mitigation, food security, biodiversity conservation and poverty alleviation.

213 Numerous examples exist of national and international policies on environment and
214 sustainability that have ignored grasslands. For example, halting and reversing ecosystem
215 degradation is a central goal of the SDGs (Goal 15), the UN Framework Convention on Climate
216 Change (UNFCCC), and the Convention on Biological Diversity (CBD), but there is no explicit
217 mention of grasslands in any of them. The SDG targets also fail to mention grasslands, whereas
218 forests, oceans and coastal ecosystems are repeatedly referenced in targets and indicators.
219 Further, while grassland regions contain around half the world’s 234 Centres of Plant Diversity
220 - sites of global botanical importance based on endemism and species richness - they were not
221 mentioned in the Aichi Biodiversity Targets of the CBD¹⁴. The European Biodiversity Strategy
222 to 2020⁴⁸ also makes little mention of grasslands, whereas forests feature heavily.

223 Pressures on grasslands have also been accentuated through the unforeseen
224 consequences of global and national environmental and socio-economic policies. In particular,
225 grasslands are increasingly targeted in carbon sequestration programs that emphasize tree
226 planting, such as REDD+ and China’s Grain-to-Green Project, which often leads to their
227 degradation. The Chinese government, for example, has invested heavily in tree planting, with
228 tree coverage of previously non-forested areas increasing at an average rate of 0.15 million
229 hectares per year over the last 25 years. However, there is major concern that tree planting in
230 arid and semi-arid grasslands with water-demanding trees is exacerbating grassland
231 degradation, reducing plant diversity, damaging soils, and increasing water shortages⁵⁰⁻⁵².
232 Also, large tracts of grassland in Brazil have been identified as targets for tree planting, posing
233 a major threat to these ancient and highly diverse ecosystems²⁵. An estimated 7.41 million ha
234 of eucalypt and 2.07 million ha of pine plantation already occupy mostly former savannahs and
235 grassland, at an astonishing rate of 0.4 million ha per year from 2013-2017⁵³. Indeed, it has
236 been calculated that meeting land restoration and protection targets would increase global tree
237 cover by 4M km², often at the expense of grasslands⁵⁴. Furthermore, models of global forest
238 restoration potential have identified natural grasslands as sites for restoration using tree
239 planting, with potential to contribute to carbon sequestration and climate mitigation⁵⁵. The

240 logic of such proposals for climate mitigation, however, is increasingly been challenged on the
241 basis of their damaging impact on natural grasslands and because their carbon sequestration
242 potential is considered inflated^{39,56}.

243 Grasslands can be more resilient carbon sinks than forests⁵⁷, and afforestation can cause
244 soil carbon loss^{58,59}, soil acidification and nutrient-depletion⁶⁰, especially when trees are
245 planted in natural grasslands⁵⁹, making them prone to carbon loss from fires⁵⁶. Large-scale
246 afforestation also leads to changes in surface albedo, given that forests absorb more short-wave
247 radiation than grasslands, thereby creating a warming effect^{60,61}. As such, changes in albedo
248 resulting from afforestation can reduce or even negate benefits of increased carbon capture,
249 potentially leading to a net warming effect of tree planting^{60,61}. Another issue is that policies
250 such as REDD+ focus primarily on carbon sequestration in aboveground tree biomass, while
251 healthy and restored grasslands can store comparable amounts of organic carbon as forests, but
252 mainly belowground^{57,63}. Grasslands have also been shown to be more effective than forests in
253 providing soil erosion control and water protection in semi-arid ecosystems⁶⁴, and the
254 conversion of grassland to forest, either through natural regeneration or afforestation, can be
255 highly detrimental to people who depend on grasslands for forage, game habitat, water
256 reserves, and cultural services^{56,65,66}.

257 The lack of emphasis on grasslands in international and national policies has a long
258 history. The 1954 Constitution of the People's Republic of China, for example, recognised
259 forests, mines, rivers, wastelands (which probably included grasslands) and other resources,
260 while grassland law, aimed at protecting, developing and making rational use of grasslands,
261 did not appear in the constitution until 1985. Grasslands in India have also been historically
262 undervalued in national policies, and even today they are widely considered to be unproductive
263 wastelands of limited value, leaving them vulnerable to land conversion^{67,68}. In Brazil, while
264 major progress has been made in the conservation of forest ecosystems, non-forest biomes,
265 including the Cerrado and Pampas grasslands, have been largely neglected, despite being
266 among the most species rich grasslands in the world⁶⁹. Similarly, and despite their ancient
267 origin, Asian savannahs have been considered to be degraded forest since colonial times, which
268 has led to inappropriate management and policies that promote their degradation and loss⁷⁰.

269 If grasslands are to be valued and managed sustainably, then both global and regional
270 policy must change to recognize the value of grasslands as providers of multiple ecosystem
271 services and hotspots for biodiversity, and to establish targets for their protection and
272 sustainable management. Approaches could comprise specific inclusion of grasslands in
273 conservation laws and policies and ensuring conservation policies do not have perverse

274 consequences for grasslands, such as promoting fire suppression or afforestation. New policies
275 to promote and fund appropriate management and restoration of grasslands are also needed,
276 including penalties for degrading activities such as overgrazing. As an example, China's
277 grassland law includes a range of policies that have been implemented to protect healthy and
278 restore degraded grasslands⁷¹.

279

280

281 ***Standardised assessment.*** The coherent evaluation of grassland condition across the World
282 requires standardised approaches. These should assess the severity of grassland degradation
283 and its consequences for biodiversity and ecosystem services and evaluate the success of
284 restoration schemes. Currently, restoration efforts are disjointed across regions and carried out
285 by a wide range of organisations, which often leads to incompatible and inaccessible datasets,
286 and a lack of communication about successful methods^{72,73}. The diversity of grasslands across
287 and within regions, and the many drivers of degradation, mean a fully uniform set of guidelines
288 is impractical. Yet, we suggest that standardised approaches are needed to underpin effective
289 decision-making.

290 Progress can be made by defining grassland degradation from a socio-ecological
291 perspective, where grassland is considered degraded if the supply of multiple ecosystem
292 services falls short of that demanded by stakeholders⁷⁴. Different stakeholder groups prioritise
293 different combinations of ecosystem services^{75,76}; as such, the definition of degradation
294 depends upon the stakeholder, and needs to be tailored accordingly (Figures 2-4). In some
295 cases, grassland degradation is apparent to all stakeholders, for instance when overgrazing
296 leads to loss of vegetation, reductions in soil organic matter and consequent soil erosion⁷⁷.
297 However, in other cases, perceptions of degradation and restoration success may differ: the
298 increase in plant production but accompanying loss of plant species diversity resulting from
299 fertilizer use might be considered an improvement by pastoralists, due to increased forage
300 production, but as degradation by conservationists concerned with biodiversity protection⁷⁸.

301 The severity of degradation and the relative benefits of different grassland restoration
302 practices can be assessed by combining measures of ecosystem service supply and stakeholder
303 priorities and demands. Such measures can show how alternative restoration options differ in
304 their impacts on ecosystem service supply and in their resulting benefits for different
305 stakeholder groups. We recommend that this is achieved via a general five-step approach for
306 standardising the assessment of grassland degradation and restoration (Figure 2). The results
307 of this approach can also be used to support negotiation over restoration management.

308 The first step is to assess the demand for, and relative priority of, different ecosystem
309 services by multiple stakeholder groups, for instance via social surveys of many stakeholders,
310 or representative community leaders^{75,39} (Figure 2). It is vital that all major stakeholder groups
311 are represented and the full range of relevant ecosystem services, including non-material
312 (cultural) benefits considered, otherwise bad management decisions may result and potential
313 conflict⁸⁰. Once ecosystem service demand is determined, indicators need to be identified and
314 measured for each of the services. We propose that these should fall into two classes: general
315 indicators, which measure ecosystem services for all stakeholder groups; and specific
316 indicators for services used by a subset of the stakeholder community. In step 2, general
317 indicators should be inexpensive and easy to measure over large scales and relate to key
318 regulating functions that underpin all other ecosystem services. Such properties could include
319 the cover and type of vegetation present, aboveground biomass, and soil properties such as
320 organic matter, nutrient content, and pH. A presentation of possible measures and the
321 ecosystem services they represent is in Table 1.

322 The next step is to find tailored indicators for the specific ecosystem services used by
323 different stakeholder groups. While the general indicators proposed above provide broadly
324 comparable information on grassland condition, perceptions of grassland degradation differ
325 between stakeholder groups⁸¹. Therefore, in addition to general indicators, other variables
326 should be identified locally for these context-specific ecosystem services. These services and
327 their indicators should be identified via stakeholder engagement, with indicators being
328 developed according to local conditions. In the French Alps, for example, surveys identified
329 that aesthetic value of mountain grasslands was essential to local farmers and tourists, and
330 related not only to flower diversity, but also to the absence of a build-up of plant litter⁸². As
331 another example, the cover of woody species can be used as an indicator of the vital regulating
332 service of shade provision in East African grasslands⁸³. Visually estimable indicators can also
333 be employed in such assessments to ensure both relevance and cost-effectiveness.

334 In the fourth step, general and specific measures can be combined to calculate an
335 integrated index of grassland condition that can be related to local environmental conditions,
336 management factors, global change drivers and restoration management. The creation of such
337 indices can be achieved by adapting approaches developed for measuring the co-delivery of
338 multiple ecosystem services based on stakeholder preferences^{74,84}, and other participatory
339 multi-criteria analyses of ecosystem services⁸⁵. Both approaches can also be used to weigh the
340 measures included in the calculation of the index, for example by encouraging representatives
341 of different stakeholder groups to assign scores to the ecosystem services considered⁸³. In the

342 final step, we suggest that conclusions from this standardised approach are used to inform
343 management decisions regarding which restoration options should be employed and where.
344 This is best resolved in participatory approaches, so that stakeholders share understanding, and
345 conflict is minimised (Figure 2, and see Shared Understanding section below). The best options
346 for whole communities can be estimated by weighting stakeholder groups equally in a
347 community level metric, or by evaluating which restoration options minimise trade-offs
348 between groups. As a single restoration practice may not benefit all stakeholder groups, the
349 compartmentalisation of the landscape into different restoration options should also be
350 considered (Figure 2).

351 The approach presented here would not only allow for more detailed examination of
352 grassland condition with regards to local needs, but also a better means of assessing the severity
353 of grassland degradation and restoration success at both local and larger scales. In figures 3
354 and 4, we demonstrate how this approach can be applied in practice and how different
355 restoration strategies create trade-offs or synergies for stakeholders in two regions: temperate
356 European grasslands and in the arid and semi-arid grasslands of Eastern Africa (Figures 3 and
357 4). We advise readers to see Supplementary Material for details on the source of these
358 estimates, and Manning et al.⁷⁴ for a tutorial and R code for calculating such metrics.

359

360 ***Shared understanding and trade-offs.*** When assessing grassland restoration options that best
361 suit different stakeholder groups, there will likely be differences in preferred options (Figures
362 3 & 4), which might cause disagreement and conflict. To address this, there is a need for shared
363 understanding, based on best knowledge and practice and operational constraints, of the
364 potential for different grassland restoration options to supply different ecosystem services^{86,87},
365 which can be quantified using the same tools as described above for assessing degradation.
366 Such an approach could help identify which restoration option might best deliver the needs of
367 all stakeholder groups, thereby supporting resolution of disagreements and conflict.

368 To illustrate these trade-offs and how they may be resolved we present two detailed
369 examples in which the standardised approach described above is applied. In both, fundamental
370 trade-offs caused by degradation and restoration options prevent the co-supply of all ecosystem
371 services. The first example is taken from European grasslands (Figure 3), where extensive
372 management promotes functions related to water and soil carbon storage, which support
373 ecosystem services linked to water and climate regulation. In contrast, intensive management
374 with fertiliser use and high grazing pressures leads to faster rates of nutrient cycling and higher
375 plant productivity, thereby promoting provisioning services^{78,88}. However, intensive

376 management also reduces plant and soil biodiversity and causes soil compaction, which
377 diminishes the supply of several ecosystem services demanded by stakeholders, leading to
378 reduced overall benefits to stakeholder groups (Figure 3). To reverse these impacts, restoration
379 to high diversity grassland can enhance the supply of multiple ecosystem services, and
380 therefore provides substantial benefits to all stakeholder groups. Despite this, the abandonment
381 of agricultural management altogether, which leads to woody plant encroachment, may be the
382 best option to provide some of the benefits prioritized by national decision makers, namely
383 water quality and climate regulation. Given these various trade-offs, and depending on
384 European or national objectives, and on regional conditions⁸⁹, collective choice would likely
385 favour a mosaic of all three grassland types with the restoration of high diversity grassland,
386 accompanied by the segregation of intensive production and climate regulation into specific
387 regions and landscapes⁹⁰.

388 The second example is taken from arid and semi-arid Eastern Africa (Figure 4), where
389 nomadic pastoralists favour provisioning services of forage production and natural products
390 from grasslands, while charcoal producers derive their livelihoods from native woody
391 vegetation and, increasingly, from invasive woody species such as *Prosopis juliflora*. Because
392 the invasion of *P. juliflora* transforms grasslands into scrubland with bare soil⁹¹, the supply of
393 most ecosystem services demanded by stakeholders is greatly reduced, but fuel wood and
394 charcoal production are promoted⁸³. The supply of other services, such as soil carbon storage,
395 is context-dependent; soil conditions are improved when *P. juliflora* invades grasslands
396 degraded by overgrazing and other forms of mismanagement. However, if undegraded
397 grassland is invaded, the supply of these services is reduced⁹² and all stakeholders, apart from
398 charcoal producers, lose ecosystem service benefits, and so perceive the invasion as
399 degradation. This example also clearly demonstrates that successful restoration depends on
400 stakeholder perspectives as all restoration options benefit the priority services of some groups,
401 while diminishing those of others. These constraints preclude a single win-win-win restoration
402 option. Therefore, regional-scale restoration that establishes a landscape mosaic, including
403 livestock grazing and patches of conserved or abandoned land, is likely to be required for all
404 three stakeholder groups to be satisfied. It should be noted, however, that such a landscape
405 mosaic would require continuous management of *P. juliflora*, otherwise it will re-invade
406 grasslands and conserved land.

407 Identifying the presence and cause of trade-offs is a first step towards management and
408 policy solutions to grassland degradation. We suggest that the approaches presented here could
409 be combined as a decision support tool in ecosystem management⁹³. This can be used as part

410 of a ‘landscape approach’ to management, which emphasises close collaboration and shared
411 understanding between stakeholder groups to identify and mitigate land management conflicts.
412 Depending on context, this approach could allow for the allocation of different restoration
413 options to different sections of the landscape, as in semi-arid grasslands of Eastern Africa
414 (Figure 4), or the identification of win-win options where a single restoration practice might
415 suit multiple stakeholder groups, such as in European grasslands (Figure 3). These approaches
416 are already considered in existing grassland management approaches, such as the Participatory
417 Rangeland Management (PRM) approach, which has been developed to improve the
418 sustainable management of rangeland resources and their security of access for local rangeland
419 users in arid and semi-arid grasslands⁹⁴. This approach includes a step where stakeholder
420 groups, through extensive meetings, discussions and negotiations, develop a rangeland
421 management agreement that considers the needs of all stakeholder groups involved. This
422 agreement should ultimately become a binding contract document between a representative
423 rangeland management institution and the appropriate government authority. The PRM was
424 initially developed by a consortium including the International Livestock Research Institute in
425 Ethiopia and has recently been applied in Kenya and Tanzania⁹⁵. A similar approach called
426 Participatory Rangeland Management Planning (PRMP) has been adopted by the International
427 Union for Conservation of Nature (IUCN)⁹⁶.

428 We note that complex trade-offs might prevent the formulation of simple solutions, and
429 in such situations multi-criteria decision analysis (MCDA) can be used to support deliberation
430 among groups and help steer them towards a compromise where local communities have a
431 voice and multiple perspectives of grassland degradation are accommodated⁹⁷⁻⁹⁸. Advanced
432 MCDA approaches also allow for the incorporation of socio-economic factors, such as access
433 rights and ease, and power relationships between stakeholder groups across scales, which will
434 determine the ultimate benefits of ecosystem services to different groups⁹⁹. However, to ensure
435 better outcomes from decision tools such as MCDA it is also important to consider optimal
436 allocation and prioritisation of limiting resources to actions¹⁰⁰, especially since funds for
437 grassland restoration are often limited.

438
439 ***Exemplars of grassland restoration.*** Evidence of success or failure of grassland restoration
440 programmes is scant^{101,102}. Accordingly, there is a need for standardised and accessible
441 reporting of restoration successes and failures, and ongoing monitoring of grassland restoration
442 programmes and their outcomes in different parts of the world. An important step for achieving
443 this is the sharing of restoration knowledge which requires the formation of networks of

444 scientists and practitioners that exchange knowledge on successful, but also unsuccessful,
445 restoration efforts (Box 1). Such networks could form a platform for case studies illustrating
446 novel ideas and approaches to restoration at local and landscape scales in different parts of the
447 world (Table 2; Figure 1). They could also illustrate the successful translation of research into
448 improved restoration and management, which should include the co-creation and co-
449 implementation of technical solutions to grassland restoration, and the political and socio-
450 economic conditions that made them possible.

451 Case studies reporting technical solutions to grassland degradation at scales of hundreds
452 to thousands of hectares are particularly scarce, especially for developing countries in semi-
453 arid and arid regions of the world. Studies summarized in Table 2, however, indicate that
454 grassland restoration is feasible, at least starting from low levels of degradation, and that
455 technical and socio-ecological solutions often need to be combined to achieve long-term
456 management goals. For example, grasslands in Inner Mongolia are challenged with a diverse
457 set of stakeholder demands, including the need for environmental services, such as increased
458 soil carbon storage, and the local herders' wish to retain traditional livelihoods while increasing
459 their income, and an increasing demand for red meat by the wider population in China and
460 elsewhere⁷¹. To address these demands of various stakeholder groups, a grassland management
461 system was developed that built on optimizing income from livestock production per unit area,
462 which was achieved by increasing animal growth rates (and thus reaching marketable size
463 earlier) by decreasing livestock numbers by approximately 50%⁷¹. Thus, the system created a
464 win-win situation by improving local herders' income and enabling the recovery of grasslands
465 and associated ecosystem services (Table 2).

466 In northern Ethiopia, several years of excluding livestock from mountain grasslands led
467 to the recovery of plant species richness, forage and wood production, and soil health, and thus
468 helped to improve the supply of services prioritized by local stakeholders¹⁰³⁻¹⁰⁵. However, the
469 uptake of technical solutions for restoring degraded grasslands depends on socio-economic
470 factors, including reconciliation with traditional land-tenure systems, well-established
471 communication pathways, and supporting institutional settings and policies (Table 2). For
472 example, implementation of sustainable grassland management in Inner Mongolia benefits
473 from recent efforts of the Chinese government to emphasize sustainable grassland management
474 in agricultural policies¹⁰⁵, while the success of the grazing enclosure programme in northern
475 Ethiopia will depend on successful communication and negotiations between local stakeholder
476 groups and policy as well as among local stakeholders to ensure land use enforcement and
477 sharing of management responsibilities and harvesting benefits¹⁰³⁻¹⁰⁵.

478
479
480 **Research innovation:** A new generation of environmental research offers the potential to
481 improve the assessment of grassland degradation and provide novel solutions to restoration. At
482 global and regional scales, remote sensing technologies can evaluate both the extent and status
483 of grasslands and inform the spatial targeting of large-scale restoration efforts. For example,
484 maps of general degradation indicators, such as primary productivity, standing biomass, soil
485 moisture, phenology, soil organic carbon¹⁰⁶⁻¹⁰⁸, could be combined with spatial information on
486 climate, edaphic and socio-economic data to identify national and global patterns of grassland
487 degradation and pin-point locations where restoration efforts may have the greatest impact.
488 While current global maps of primary productivity and biomass carbon still need improvement,
489 they offer a means to monitor the dynamics of grassland aboveground biomass and evaluate
490 the degradation of global grasslands^{109,110}.

491 In regions where semi-natural grassland remnants are fragmented and persist within a
492 matrix of forested or intensive agriculture land uses, new research could facilitate landscape
493 and regional scale restoration of inter-connected grasslands. There is increasing evidence that
494 grassland dynamics and function are enhanced by connectivity that encourages flows of species
495 and biogeochemical processes^{111,112}. In theory, such spatial networks have improved resilience
496 to perturbations, such as climate extremes, and the ability to adapt to a changing
497 environment¹¹³. Research is needed to apply these emerging concepts to landscape design of
498 restoration that allows ecological flows of species and processes^{114,115}, although at a practical
499 level, there is already considerable opportunity to trial these approaches in large-scale
500 restoration programmes¹¹⁶. Such approaches also require improved understanding of the tele-
501 coupling of different landscape elements, whereby human-induced processes in one area
502 impact on distant areas. For example, pastoralists in arid regions depend on grazing areas along
503 major water bodies, which are also of interest to other stakeholders for crop production or
504 biodiversity conservation¹¹⁷. Loss of such grazing areas forces pastoralists to overgraze rain-
505 fed grasslands elsewhere, leading grassland degradation¹¹⁸.

506 Another fast-developing area of ecological research that could assist grassland
507 restoration programmes concerns the application of knowledge of aboveground-belowground
508 interactions^{119,120}. A wealth of studies show that high grassland plant diversity can enhance
509 multiple ecosystem functions, both above- and belowground, and increase the resistance of
510 plant production to environmental perturbations, such as climate extremes¹²¹⁻¹²³; such
511 knowledge could be applied to degraded grassland to restore ecosystem functions and their

512 resilience to environmental change^{124,125}. Benefits can also be realised through informed
513 selection of plant species based on functional traits, especially root traits, which has been
514 shown to be an effective way of enhancing grassland multifunctionality¹²⁶ and the physical
515 properties of degraded soils¹²⁷. Research points to the potential of soil inoculation and plant
516 translocation from donor sites to enhance the recovery of degraded ecosystems and steer plant
517 community development^{128,129}. There is also increasing awareness that biodiversity at multiple
518 trophic levels is critical to developing multifunctional grassland ecosystems¹³⁰. Studies show,
519 for instance, that diversifying livestock promotes plant diversity and grassland
520 multifunctionality¹³¹, especially at moderate levels of grazing¹³², and multi-trophic restoration
521 might be implemented via improved understanding of interaction networks and how
522 interventions can facilitate robust networks in restored ecosystems¹³³.

523 A potential hurdle to the adoption of new research innovation is that the majority of
524 scientific knowledge underpinning grassland restoration comes from studies done at a local
525 scale and in a handful of regions, especially North America, Australasia and Europe. As such,
526 there is a need for new, long-term research to underpin approaches to large-scale restoration of
527 degraded grasslands, especially in Asia, Africa and South America where grassland
528 degradation is widespread¹³⁴. One solution to these geographic gaps might be the use of meta-
529 level analyses¹³⁵⁻¹³⁶ and data collected using standardised methodology from globally
530 distributed grassland studies such as the Nutrient Network¹³⁷ and HerbDivNet¹³⁸. These can
531 help to identify generalisable impacts of biotic and environmental drivers and disturbances,
532 such as grazing and nutrient enrichment, on grassland diversity and function, that may be
533 transferred to regions where data is lacking. It should be noted, however, that large gaps in
534 environmental and management conditions still need to be filled, and the ecological context
535 will affect the specific applicability of any generalisations¹³⁹. At all scales, new technology and
536 practices need to be embedded within socio-ecological approaches to grassland restoration to
537 ensure empowerment of local people and best outcomes for their quality of life.

538

539 **Summary and future perspectives**

540

541 Despite progress in understanding the causes and consequences of grassland degradation, and
542 in developing restoration techniques, there remain many barriers to halting grassland
543 degradation and effective restoration. Here, we have argued that overcoming these barriers
544 requires an integrated socio-ecological approach to grassland degradation and restoration, that

545 not only demands greater recognition of grasslands in global policy, but also standardised
546 approaches for assessing grassland degradation and restoration success. It also requires the
547 adoption of recent innovations in environmental sciences for both detecting grassland
548 degradation and enhancing restoration success, along with enhanced knowledge transfer and
549 data sharing regarding restoration experiences.

550 Many examples exist of national and international sustainability policies that have
551 ignored grasslands, including the SDGs and CBD. Giving due attention in sustainability policy
552 to grasslands and the ecosystem services they provide, on a par with other biomes such as
553 forests, is therefore essential for the future protection of healthy grasslands and restoration of
554 those that are degraded. The UN Decade on Ecosystem Restoration (2021-2030) and CBD
555 post-2020 global biodiversity framework provides an opportunity to set future targets for the
556 protection, sustainable management and restoration of grasslands, and prevent damaging
557 practices such as the planting of trees in natural grasslands¹⁴⁰. Put simply, if grasslands are to
558 be managed sustainably, then both global and regional policy must be revised to recognize the
559 value of grasslands for multiple ecosystem services and establish targets for their protection,
560 restoration and sustainable management. There is hope in some recent initiatives. The
561 Worldwide Fund for Nature has recently launched a ‘Global Grassland & Savannah Dialogue
562 Platform’, with the aim to “develop consensus around human and biological importance of
563 these ecosystems”. The UN Decade on Ecosystem Restoration has recently added ‘grasslands,
564 shrubland and savannahs’ to their set of focal ecosystems. Furthermore, while the Aichi targets
565 of the CBD largely ignored grasslands, and the draft of its post-2020 framework lacks
566 specificity, the most recent synthesis of the proposals of the parties includes mentions of
567 grasslands (albeit other ecosystems such as forest and wetlands are discussed more frequently).

568 Whilst we demonstrate how our five-step approach can be applied using specific case
569 studies, future research is needed to test this approach in different contexts and at local and
570 larger scales. Research is also needed to develop and test promising new ways of assessing
571 grassland degradation and restoration, and to harness ecological knowledge for restoration
572 success. This will require ambitious, interdisciplinary national and international research
573 programmes of the kind that could be facilitated by major research and innovation schemes
574 targeted at achieving Sustainable Development Goals. We hope the approach we present
575 provides a basis for such future research aimed at the assessment of grassland degradation and
576 restoration in the context of stakeholder needs.

577 The actions we raise for developing effective socio-ecological solutions to the
578 degradation and restoration of global grasslands are not exhaustive and many challenges

579 remain. Nevertheless, we hope that they provide a guide to future research and policy needs
 580 for halting grassland degradation and achieving restoration success. We also hope that they
 581 serve to raise awareness of the plight of global grasslands and the need for urgent action to halt
 582 grassland degradation and enhance restoration success, thereby conserving the many socio-
 583 economic, cultural and ecological benefits that grasslands provide.

584

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972 **Box 1: Platforms for effective grassland restoration**

973 Examples of platforms that provide information on successes and failures of grassland
974 restoration programmes are beginning to emerge. The Society for Ecological Restoration
975 (SER) recently launched their Restoration Resource Centre, an online, publicly accessible
976 platform for exchanging knowledge and experience through ecological restoration projects,
977 publications, and other resources from around the world. The World Overview of Conservation
978 Approaches and Technologies (WOCAT) database provides a format and a platform for
979 knowledge exchange regarding Sustainable Land Management (SLM) technologies and
980 approaches and is recognized as the primary recommended database by UNCCD. The Global
981 Landscape Forum (GLF), led by the Centre for International Forestry Research (CIFOR) and
982 supported by the UN and World Bank, provides a mobile platform for discussions leading to
983 action on sustainable land use, including land restoration. Grasslands are currently
984 underrepresented within these databases, which are still under development and continually
985 being expanded. Nevertheless, they demonstrate the potential for user-driven platforms to
986 provide valuable insight into effective restoration practices in different parts of the world. To
987 be effective, such platforms should include a wide range of restoration programmes, both local
988 and regional, and small and large scale. They should also include less successful programmes
989 and identify why they have not reached their objectives. By comparing case studies, it is
990 possible to understand how the effectiveness of specific restoration methods differ
991 geographically and depending on the degree of degradation and socio-ecological context, the
992 first step towards finding general rules, both social and ecological, for successful restoration.

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995 **Figure legends**

996

997 **Figure 1.** The extent of degraded grasslands worldwide with examples of paired non-
 998 degraded (left) and degraded (right) grasslands from different locations around the world.
 999 Grassland classification follows the UN FAO Land Cover Classification System (LCCS)
 1000 (data downloaded at <https://lcviewer.vito.be/2015>¹⁴¹ with tundra ecosystems excluded.
 1001 Degradation is measured as greenness changes, as measured by rain use efficiency (RUE)
 1002 adjusted Sum Normalized Differential Vegetation Index (NDVI) between (1981-2015)¹⁴²
 1003 with regions showing a reduction in greenness of 0.01 being classed as degraded. Therefore,
 1004 much degradation involving vegetation change is not shown. Degradation is caused by many
 1005 factors, including overgrazing, fertilization, tree planting, and invasive species. Image credits:
 1006 United States (L. Brudvig), United Kingdom (L. Hulmes), Qinghai-Tibet Plateau, China (R.
 1007 Bardgett), Inner Mongolia (R. Bardgett), Brazil (G. Durigan), Kenya (R. Bardgett), India (M.
 1008 Sankaran, S.K. Chengappa), and Australia (S. Prober).

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 1010
 1011 **Figure 2.** Standardised assessment of grassland degradation and restoration. This conceptual
 1012 figure shows the steps required to assess grassland condition and select restoration options. The
 1013 relationship between this approach and the other strategies described in this article are shown
 1014 on the side panel. For specific and quantitative examples of these steps, see Figures 3 and 4.

1015
 1016 **Figure 3.** Assessing ecosystem service trade-offs in degraded and restored European
 1017 grasslands. In this system degradation caused by intensive agriculture reduces plant and soil
 1018 biodiversity and causes soil compaction. This diminishes the supply (SD) of ecosystem
 1019 services demanded (Di) by stakeholders, resulting in a low multifunctionality (MF) for all
 1020 stakeholder groups (Di x SD). Restoration to high diversity, perennial grassland (SR1:
 1021 Perennial grassland) enhances the supply of multiple ecosystem services, though only
 1022 moderately for fodder production and climate regulation through carbon storage. Thus, among
 1023 restoration options it provides the greatest multifunctionality to all stakeholder groups.
 1024 Restored perennial grassland also provides the best option for priority ecosystem services (PF)
 1025 for conservationists, but also for local farmers given their demand for soil fertility and cultural
 1026 identity, which compensates for lower fodder production than intensive grassland. However,
 1027 abandonment of agricultural management may be seen as favourable by some national decision
 1028 makers due to their prioritization of water quality and climate regulation.

1029
 1030 **Figure 4.** Assessing ecosystem service trade-offs in degraded and restored grasslands in arid
 1031 and semi-arid Eastern Africa. In these regions, invasion of alien woody plants such as *Prosopis*

1032 *juliflora* transforms grasslands into scrubland with bare soil. While the supply (SD) of most
1033 ecosystem services demanded (D2) by stakeholders is greatly reduced by this invasion, fuel
1034 wood and charcoal production are promoted. The supply of other services, such as soil carbon
1035 storage, is context-dependent; if grassland previously degraded by overgrazing and other forms
1036 of mismanagement is invaded, then soil conditions are improved. However, if healthy
1037 grassland is invaded, the supply of these services is reduced. All stakeholders, apart from
1038 charcoal producers, lose ecosystem service benefits (D2 x SD) if undegraded grasslands are
1039 invaded by *P. juliflora*, and so perceive it as degradation. Restoration options have differing
1040 impacts on stakeholder groups. Sustainable livestock grazing (SR2) requires fencing off or
1041 access restriction but promotes ES supply more than conservation (SR2) and supports the
1042 greatest multifunctionality (MF) from the perspective of all the considered groups.
1043 Abandonment does not restore the supply of any ES and even incurs ES losses (SR3), because
1044 *P. juliflora* consumes significant amounts of groundwater. However, considering economic
1045 priorities towards only a priority ES (PF) of fodder for pastoralists, wood products for charcoal
1046 producers and ecotourism for tourists would lead these stakeholders to respectively favour
1047 livestock production, abandonment or conservation as best restoration options (highest
1048 respective PF scores).

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1073 **Table 1.** Common ecosystem services of grasslands and indicators of associated ecosystem
 1074 structure and function parameters
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Ecosystem services	Ecosystem structure / function	Common indicators	Scale and means of measurement
Forage production (quantity)	Annual aboveground biomass production	Net primary productivity (NPP), aboveground biomass	NDVI (remote sensing), landscape scale Direct harvesting, local scale
Forage production (quality)	Protein content and digestibility	Leaf N content Presence of species of particular nutritional importance (e.g. legumes)	Remote sensing at a landscape scale Direct measures at a local scale
Forage reliability	Interannual variation in aboveground biomass production	Monitoring of NPP species composition	Remote sensing at landscape scale
Other grassland products (e.g. medicinal, food, hunting) Biofuels	Species of particular interest Woody species of interest Grass species of particular interest (e.g. flammable and high yielding species)	Species presence / abundance Cover or biomass of species of interest	Field survey at local scale Direct field measures or remote sensing
Species of cultural value	Presence of species of cultural interest	Species presence	Direct measures and local records (e.g. (National monitoring schemes and local surveys) at local scale or regional survey for large areas
Aesthetic value	Plant community composition and phenology	Flower cover, flower colour, and presence of 'unattractive' species Flowering phenology	Direct survey Phenology can be derived from species list and external databases
Biodiversity conservation value	Plant and animal (vertebrate / invertebrate) species	Presence and abundance of species of conservation value	Global Biodiversity Information Facility (GBIF) and national monitoring schemes at larger scales

Regulation of invasive exotics and other undesired species	Invasive exotic species, species of negative pastoral or cultural value	Presence / abundance of undesired species	Direct survey at local scale GBIF and national monitoring schemes at larger scales
Global climate regulation	Carbon stocks and carbon cycling processes	Soil respiration and carbon stocks Woody species biomass and vegetation carbon stocks Litter mass / depth	Direct survey at local scale Remote sensing for aboveground stocks Soil sampling and interpolation for belowground carbon stocks and fluxes
Maintenance of soil fertility	Nutrient stocks and nutrient cycling processes	Soil nutrient and carbon content Litter mass Soil enzyme activities	Soil sampling and interpolation to required scales
Maintenance of soil stability and regulation of erosion	Soil stability in the root profile Erosive flows	Evidence for erosion, bare ground cover and soil organic matter Soil aggregate stability and bulk density Plant rooting profile Direct measures of soil loss / erosive flows Water holding capacity	Field observations or remote sensing to detect erosion Direct sampling and measurement of multiple indicators
Regulation of hydrological flows	Soil water retention and flows	Soil texture and bulk density, and soil organic matter content Soil electrical and hydraulic conductivity (rapid assessment methods)	Direct survey and interpolation
Regulation of water quality	Retention and transformation of pollutants in soil	Soil physicochemical properties (e.g. texture, pH, CEC) Salinity and water table depth Soil organic matter content and available water capacity.	Direct sampling of soil and water

Nutrient and
pollutant
concentrations in
freshwater bodies

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